



Risks to biodiversity and coastal livelihoods from artisanal elasmobranch fisheries in a Least Developed Country: the Gambia (West Africa)

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Abstract

Developing nations in tropical regions harbour rich biological resources on which humans depend for food, income and employment, yet data to aid their management is often lacking. In West Africa, the diversity and fisheries of elasmobranchs are poorly documented, despite them being known to be economically important and vulnerable to overexploitation. Rapid qualitative surveys of fish processing and landing sites in The Gambia from 2010–2018 revealed valuable new data on species composition, biology, relative abundance, fisheries and utilisation by humans. Diversity and abundance was dominated by batoids, with a major component comprising a large guitarfish (*Glaucostegus cemiculus*) that was apparently targeted, and a small whiplay (*Fontitrygon margaritella*). Nearly all taxa recorded are classified by the IUCN *Red List* as Critically Endangered (angel sharks *Squatina* spp.), Endangered, Data Deficient, or Not Evaluated; several were endemic, of exceptional evolutionary distinctness, cryptic, possibly undescribed, and rare (including stingray *Hypanus rudis* not apparently recorded since description in 1870). Significant threats to biodiversity, coastal livelihoods and possibly food security are identified based on the apparent importance of elasmobranch fisheries and processing; the known inability of key taxa to withstand intensive fisheries; ‘fishing down the food web’ by intensive utilisation of *F. margaritella*; and the absence or rarity of previously common elasmobranch species that may be severely depleted in the region. This study provides data that may act as a starting point to aid sustainability accreditation of local fin-fisheries, and demonstrates the value of inexpensive and low-resolution data collection in developing countries.

Keywords Elasmobranchii · Batoid · Sharks · Rays · Bycatch · Conservation

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30 Introduction

31 Much tropical biodiversity is located in developing nations, where the threat of overex-
32 ploitation is greatest (Myers et al. 2000). Least Developed Countries (LDCs) are ‘low-
33 income countries confronting severe structural impediments to sustainable development’
34 (UNDESA 2018). In West Africa, the adjoining LDCs of Mauritania, Senegal, The Gam-
35 bia, Guinea Bissau, Guinea, Sierra Leone, and Liberia span a vast area and over 2000 km
36 of continuous coastline. Unsustainable resource exploitation in the region is reflected in
37 marked declines of terrestrial and freshwater biodiversity, while overfishing of marine spe-
38 cies threatens the food security and coastal livelihoods of up to 400 million people (Mallon
39 et al. 2015; Polidoro et al. 2016). Artisanal marine fisheries in the region are recognised as
40 a priority in terms of overexploitation, dependency of human communities, and need for
41 better data and management (Belhabib et al. 2018).

42 The Gambia is the smallest country on mainland Africa, entirely surrounded by Senegal
43 except for its coast on the eastern Atlantic. The country has a human development index
44 ranking of 160 (out of 179 nations), and food security is a major issue confronting the
45 country’s population, which is highly dependent on rice imports and therefore vulnerable
46 to fluctuations in global markets (Moseley et al. 2010; FAO 2018a). Fish is the main sup-
47 plier of animal protein in the diets of most Gambians, while fisheries and related activities
48 are the main source of income for coastal fishing communities and support the livelihoods
49 of an estimated 200,000 people (United Nations 2014).

50 Chondrichthyan fishes comprising sharks, rays (batoids) and chimaeras have been iden-
51 tified as having a higher extinction risk than most other vertebrates (Dulvy et al. 2014).
52 Limiting life characters such as relatively large size, late maturity and few young are unable
53 to withstand intensive targeted and bycatch fisheries, and five of the seven most threatened
54 families are batoids found in warm shallow coastal waters (Dulvy et al. 2014). Regional
55 extinctions of sawfishes on a global scale are now relatively well documented (Thorson
56 1982; Robillard and Séret 2006; Moore 2015; Leeney and Downing 2016), while severe
57 declines of other batoid groups such as guitarfishes have occurred but are less well known
58 (Moore 2017).

59 The West Africa region is home to a specialised elasmobranch industry of fisheries,
60 processing and trade, which was started in The Gambia by Ghanaian immigrants (Walker
61 et al. 2005; Diop and Dossa 2011). Furthermore, West Africa also has been identified as
62 one of five global hotspots to prioritize for conservation of chondrichthyans based on spe-
63 cies richness, endemism and evolutionary distinctness, and as a priority region for the con-
64 servation of Critically Endangered angel sharks (Gordon et al. 2017; Stein et al. 2018).
65 Despite this importance, little is known about elasmobranch biodiversity and fisheries in
66 West Africa, although there has been some work focusing on reproductive biology of a few
67 species (Capapé et al. 2002; Seck et al. 2004; Valadou et al. 2006). Prompted by the 2001
68 launch of a Plan of Action for the conservation and management of shark populations by a
69 regional fishery organisation (the Sub-Regional Fisheries Commission (SFRC/CSRP) elas-
70 mobranch surveys were undertaken at markets and landing sites in SFRC member states
71 countries (Mauritania, Senegal, The Gambia, Guinea Bissau, Guinea, Sierra Leone and
72 Cape Verde). These provide the only regional overview of elasmobranch fisheries (Diop
73 and Dossa 2011). Although these authors provided relative abundance of elasmobranch
74 species by country, they did not provide detailed country-specific data, and data derived
75 from fisheries observers and officers in the report may be of questionable reliability (BS,
76 pers. obs.). An accurate assessment of trends is also made difficult by landings data that

are erratic and likely to be unreliable. Between 1990 and 2016 Gambia reported an average of 911 t of sharks and rays annually (3.1% of total marine fish production, MFP), but this varied wildly from 194 to 4,022 t (and 1 to 13.2% of MFP) (FAO 2018b) and was notably different to other reports (Saine 2011). Studies of artisanal fish processing in Gambia suggest that smoked and dried elasmobranchs are of economic importance to coastal communities, mostly for export and particularly to Ghana, but also for some local consumption (Mbenga 1996; Njai 2000; Saine 2011). Very limited useful data is available on elasmobranch population trends in the Gambia, but declines in shark fisheries between 2001 and 2011 have been noted (Saine 2011), as has the severe decline of sawfishes since the mid-1970s (Leeney and Downing 2016). West Africa has recently been recognised as a global priority in terms of data collection and management of chondrichthyan fisheries (Dulvy et al. 2017).

Given the apparent importance of the region to elasmobranch biodiversity, fisheries and the people reliant on them, data on aspects such as species composition, relative abundance, and utilisation are essential in developing strategies for biodiversity conservation, fisheries management and food security. While a suite of high-resolution data would be ideal, obtaining these are not feasible in developing countries or LDCs with no, or very limited, local resources. In these situations, inexpensive techniques such as interviews and community-based monitoring can provide essential data on poorly-known marine vertebrate groups of economic and/or conservation importance, in Africa and elsewhere (Aylesworth et al. 2017; Humber et al. 2017; Braulik et al. 2018). For elasmobranchs in the tropics and subtropics, surveys of fish markets, landing and processing sites have provided a relatively inexpensive source of invaluable data on biodiversity and fisheries (e.g. White and Dharmadi 2007), although surveys in sub-Saharan Africa are, to date, very limited (Diop and Dossa 2011; Barrowclift et al. 2017). With the aim of addressing the significant current data gaps on biodiversity, relative abundance and utilisation, here we provide results of rapid, low-resolution surveys of elasmobranch landing and processing sites in The Gambia.

Methods and materials

Study area

The Gambia's physical environment is dominated by the River Gambia and its estuary, which joins the exposed coast of the tropical eastern Atlantic (Fig. 1). A long dry season extends from November to May, with a rainy season from June/July to October. The study was undertaken at fisheries landing sites and associated processing communities on the coast (Fig. 1). Artisanal fisheries are the most important of the sector in the country, and these are conducted by traditional open wooden pirogues (approximately 10–15 m length) with outboard motors, launched from the beach. Major fisheries are based on the use of encircling nets for small pelagics (e.g. bonga shad *Ethmalosa fimbriata* and *Sardinella* spp.), and set bottom gillnets for a range of demersal taxa such as sole (Soleidae) and tonguefishes (Cynoglossidae) (FAO 2007; UN 2014, unpublished data). Landing sites often have adjacent areas for processing of fish for consumption or export, such as air-drying racks and smokehouses (FAO 2007; United Nations 2014; unpublished data). Initial surveys identified Ghanatown and Gunjur as having a significant elasmobranch component in landings and/or processing areas, so later surveys focused on these locations.



Fig. 1 Map of The Gambia showing sampling locations mentioned in the text

121 Surveys

122 Surveys were conducted annually in the dry season (March/April) of 2011–2018, and also
 123 in the wet season (July) of 2010. Surveys typically comprised small groups of undergradu-
 124 ate students visiting sites once every few days over a two-week period, photographing all
 125 elasmobranchs for later collation and review. A more detailed survey was undertaken in
 126 the dry season of 2014 (30th March–11th April) when the students were joined by one of
 127 the authors (ABMM) to verify identifications and collect further information. From 2014
 128 onwards, the surveys were assisted by a local interpreter from the Ministry of Fisheries.

129 Identification, enumeration and measurement

130 Field identifications were initially made with Séret (2006), supplemented by Compagno
 131 and Roberts (1984) to differentiate similar-sized *Fontitrygon margarita* and *F. margari-*
 132 *tella*; all batoid identifications and nomenclature were later confirmed against Last et al.
 133 (2016a). Two groups of batoids, cownose rays (Rhinopteridae) and butterfly rays (Gymnu-
 134 ridae) were usually only identified to genus level because of their cryptic nature combined
 135 with dessication and damage.

Detailed enumeration, measurement and sexing of a large proportion of elasmobranchs seen was not feasible, as they were often present as large stacked piles of dried specimens ready for sale or export (e.g. at Gunjur). Qualitative and semi-quantitative observations were recorded, and subsamples of more accessible individuals (e.g. while on drying racks, or landed fresh) were measured. For intact sharks, guitarfishes, skates and electric rays total length (TL) was measured, with disk width (DW) measured for other batoids. Measurement of intact blackchin guitarfish *Glaucostegus cemiculus* was usually not possible as these were butchered into four or more sections by fishers immediately upon landing for drying purposes. To estimate size, intact head sections allowed measurement of pre-orbital length (POL; snout tip to perpendicular line drawn between anterior margin of orbits) in the field. The TL was then estimated (ETL) by multiplying POL by a factor (6) based on POL averaging about 16.6% in published studies (Ben Souissi et al. 2007 (inc. Figure 4a); Séret et al. 2016a) and 25 individuals where both TL and POL was measured in the Gambia in March 2018.

Results and discussion

Overview

At least 27 elasmobranch species were recorded, comprising 9 sharks and 18 batoids; the most important family were stingrays (Dasyatidae) with 7 species (Table 1). Batoids also dominated abundance and biomass. Of the species regularly occurring in significant quantity the most abundant was the pearl whipray *Fonitrygon margaritella* (Dasyatidae) and the blackchin guitarfish *Glaucostegus cemiculus* (Glaucostegidae). Common guitarfish *Rhinobatos rhinobatos* (Rhinobatidae) occurred regularly throughout surveys but in smaller numbers, and cownose ray *Rhinoptera* sp. (Rhinopteridae) occurred in significant quantity once. Most other species occurred occasionally or rarely.

Key commercial species

Glaucostegus cemiculus was commonly recorded at the landing and processing site of Ghanatown in all years, including the wet season survey of July 2010. It appeared that a targeted fishery operated for this species at this site, where it was often observed as the dominant elasmobranch being landed fresh and butchered into sections, or drying on nearby racks (Fig. 2a). Dorsal and upper caudal fins were always removed by fishers, apparently before or just after landing (Fig. 2b; also see Moore 2017 Fig. 1); the rostral cartilage of individual fish was also cut by fishers with a distinctive mark, presumably to identify ownership on drying racks. A total of 314 head sections measured at Ghanatown on two separate occasions in March and April 2014 provided ETL ranging from 102–204 cm, with 141–150 cm ETL the most frequently occurring size class (Fig. 3). A total of 26 whole fresh individuals at Ghanatown on 27th March 2018 comprised twenty females (128.6–240.1 cm TL, mean 181.4 cm TL) and five males (153.1–195.6 cm TL, mean 167.6 cm TL) (Fig. 2b).

The maximum TL recorded in the present study (240.1 cm) was slightly smaller than the maximum TL of 245 cm recorded from a smaller sample size (n = 79) of landings into Senegal in 1994–2000 (Seck et al. 2004) and smaller than the maximum reported TL of 265 cm (Séret et al. 2016a). A significant decline in the average size of this

Table 1 Details of elasmobranch taxa recorded during surveys of fish landing and processing sites in The Gambia, 2011–2017

Taxon	Species	Notes (F: fresh, D: dried, TL: total length, DW: disk width)	IUCN RL
Sharks			
Squalidae	<i>Squalus</i> spp.	~ 10 F mixed <i>Squalus</i> , not sexed, ~ 60–80 cm TL, 4/4/11, Banjul	NE
Squatinae	<i>Squatina aculeata</i>	1 F whole, ♀, 146 cm TL, 25/3/18, Ghanatown	CR
	<i>Squatina oculata</i>	1 D, not sexed, ~ 100 cm TL, 5/4/17, location not recorded	CR
Scyliorhinidae	<i>Scyliorhinus cervigoni</i>	2 F skinned, not sexed, ~ 60–80 cm TL, 18/4/15	DD
Leptochariidae	<i>Leptocharias smithii</i>	1 F whole, ♂ calcified claspers, ~ 50 cm TL, 13/4/16	NT
Triakidae	<i>Mustelus mustelus</i>	2 F whole, ~ 100 cm TL, 5/4/17, location not recorded	VU
Carcharhinidae	<i>Carcharhinus leucas</i>	Several split D sections, ~ 1.5–2 m TL, 5/4/2017, location not recorded	NT
	<i>Rhizoprionodon acutus</i>	See text; inc 1 F whole ♀, pup emergent, 116 cm TL, Brufut, 7/4/14	LC
Sphyrnidae	<i>Sphyrna lewini</i>	Occasional single F neonate & young (< 100 cm TL) and on drying racks in July 2010, and in March–April 2013–2015	EN
Batoids			
Rhinobatidae	<i>Rhinobatos rhinobatos</i>	See text. 26 F whole individuals examined comprised 24 ♀, 61–87 cm TL, 2 ♂, 52–56 cm TL, 1/4/14 (Gunjur) 8/4/14 (Sanyang)	EN
Glaucostegeidae	<i>Glaucosteagus cemiculus</i>	See text and Figs. 2 and 3	EN
Zanobatidae	<i>Zanobatus schoenleinii</i>	See text	DD
Torpedinidae	<i>Torpedo bauchotae</i>	1 F whole, ♀, 47 cm TL, 1/4/14, Sanyang	DD
	<i>Torpedo mackayana</i>	1 F whole, ♂ uncalcified claspers, 19 cm TL, 8/4/14, Sanyang	DD
Rajidae	<i>Raja cf. miraleteus</i>	1 F whole, ♂ mature, not measured, 7/4/14, Brufut	NE
	<i>Raja parva</i>	2 F whole: unsexed, 29 cm TL, 1/4/14, Kartong; ♂ mature, 49.6 cm TL, 8/4/14, Sanyang	NE
Gymnuridae	<i>Gymnura altavela</i>	See text. Inc. F whole individual of ~ 150 cm DW, April 2013	VU
	<i>Gymnura sereti</i>	See text. Inc. 3 F whole, unsexed, 80–100 cm DW, March 2018, Ghanatown	NE

Table 1 (continued)

Taxon	Species	Notes (F: fresh, D: dried, TL: total length, DW: disk width)	IUCN RL
Dasyatidae	<i>Bathytoshia lata</i>	1 F whole, unsexed, ~ 60 cm DW, 2/4/11, Ghanatown	LC
	<i>Dasyatis marmorata</i>	2 F whole, unsexed, ~ 30 cm DW, 17/4/15 & 24/3/2018, location not recorded	DD
	<i>Dasyatis pastinaca</i> complex	1 F whole, unsexed, ~ 50 cm DW; 1 F whole, ♀, ~80 cm DW, 18/4/15, location not recorded; 1 D, ~30 cm DW 21/4/15, probably Ghanatown	DD
	<i>Fontitrygon margarita</i>	~ 10 F whole, unsexed, ~ 50 cm DW, 1–2/4/11; 2 F whole, unsexed, ~ 50 cm DW 30/3/14	EN
	<i>Fontitrygon margaritella</i>	See text and Figs. 4 and 5	DD
Myliobatidae	<i>Hypanus rudis</i>	1 F whole, ♀, 80 cm DW, 2/3/11; 2 D, unsexed, ~ 60 cm DW, 2014 and 2015 (all Ghanatown)	DD
	<i>Taeniriuops grabatus</i>	1 F whole, ♀, 111 cm DW, 26/3/18, Ghanatown	DD
	<i>Aetomylaeus bovinus</i>	1 F whole, ♀, ~75 cm DW, 2/4/11, Ghanatown; 1 F whole, ♂ mature, ~ 120 cm DW, 24/3/18, Ghanatown	DD
	<i>Rhinoptera</i> sp. indet.	See text and Fig. 7. 961 unsexed D individuals inc. ?embryo of 22 cm DW, 32–85 cm DW	–

IUCN RL IUCN red list of threatened species, *CR* categories are critically endangered, *EN* endangered, *VU* vulnerable, *NT* near threatened, *LC* least concern, *DD* data deficient, *NE* not evaluated



Fig. 2 Blackchin guitarfish *Glaucostegus cemiculus* being landed and processed in The Gambia. **a** Sections on drying racks adjacent to Ghanatown beach landing site, April 2014 **b** whole animals with dorsal and caudal fins removed, Ghanatown, 27th March 2018 (credit: Jess Holdsworth)

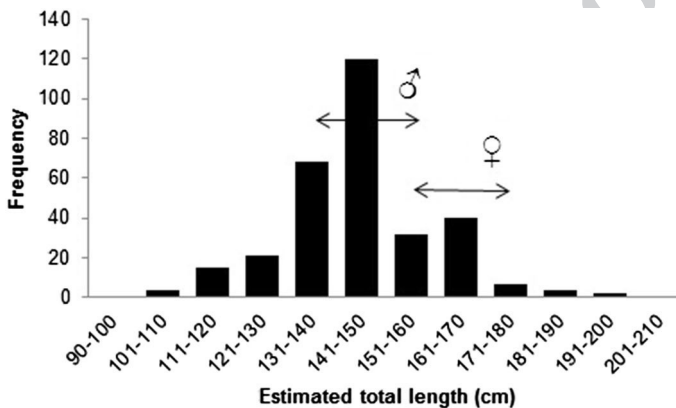


Fig. 3 Size-frequency (Estimated Total Length; see “Methods”) distribution of blackchin guitarfish *Glaucostegus cemiculus* on 31st March and 11th April 2014 at Ghanatown, The Gambia. Arrows indicate approximate size by which most or all males and females are mature, based on other studies of this species in West Africa (Seck et al. 2004; Valadou et al. 2006; Séret et al. 2016a)

species in Mauritania was reported between 1998 and 2007 (Diop and Dossa 2011). The size at maturity for *G. cemiculus* has been reported as ranging between 153 and 174 cm TL for females and 138 and 154 cm TL for males (Seck et al. 2004; Valadou et al. 2006; Séret 2016a). Based on this, many of the (unsexed) individuals in the present study were likely to be immature (Fig. 3), consistent with findings from a larger study in Mauritania (Valadou et al. 2006).

Glaucostegus cemiculus has previously been reported as frequent or abundant in much of the region (Diop and Dossa 2011). The Red List assessment for *G. cemiculus* is Endangered based on biological vulnerability (e.g. low fecundity), exposure to intensive fisheries throughout its range, high value, and evidence of declines (Notarbartolo di Sciara et al. 2016; based on a 2007 assessment). These declines have included extinction in the northern Mediterranean, a decrease in abundance in Guinea-Bissau, and an



Fig. 4 Typical examples of pearl whiplay *Fontitrygon margaritella* forming a major component of elasmobranch processing at Gunjur, The Gambia. **a** Piles of dried individuals, presumably for onward transport or sale (1st April 2014) **b** Drying individuals (4th April 2014)

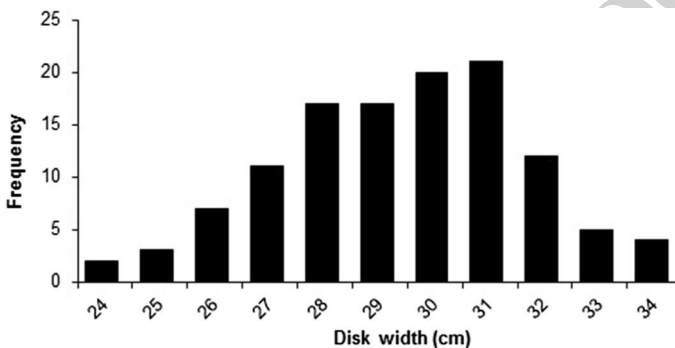


Fig. 5 Disk width-frequency distribution of a subsample of pearl whiplay *Fontitrygon margaritella* Ghanatown, The Gambia, 7th April 2014 (all individuals female excepting one 24 cm DW male)

observed decrease in the size of landed individuals from adults to juveniles in a targeted fishery in Mauritania between 1998 and 2003 (Notarbartolo di Sciarra et al. 2016).

The pearl whiplay *Fontitrygon margaritella* was commonly recorded in all years and sometimes extremely abundant at processing sites, indicating intensive utilisation. As a typical example, at Gunjur on 1–4 April 2014 several piles of hundreds of dried elasmobranchs consisted mostly of *F. margaritella* (Fig. 4a) were recorded, along with several drying racks totalling an estimated 2–3000 elasmobranchs, almost entirely *F. margaritella* (Fig. 4b); the latter included numerous small (ca. 10 cm DW) individuals. A measured subsample of 119 drying *F. margaritella* on a different occasion (Ghanatown, 7th April 2014) all but one of which were female, showed a modal and maximum DW of 31 and 34 cm respectively (Fig. 5), expanding the known size of this species (30 cm, Séret 2016b). In marked contrast to our findings, this species (as *D. margaritella*) was not recorded at all in previous surveys of the region except in Guinea where it was ‘very rare’ (Diop and Dossa 2011). The current Red List assessment for *F. margaritella* is Data Deficient (Marshall and Cronin 2016, accessed in 2007) as knowledge of fisheries interactions at the time was uncertain, and it was thought that its small size may minimise targeted capture. Our data suggest that, targeted or otherwise, *F. margaritella* is subject to intensive and

sustained fisheries mortality at all life history stages in The Gambia and in urgent need of re-assessment.

Critically endangered species

A single individual each of two Critically Endangered angel shark species were recorded. In April 2017 a smoothback angel shark *Squatina oculata* of an estimated 100 cm TL was found split and dried with pectoral, dorsal and caudal fins removed (Table 1, Fig. 6a); in March 2018, a single whole female of the sawback angelshark *Squatina aculeata* of an estimated 150 cm TL was recorded, which also had dorsal and caudal fins removed (Table 1, Fig. 6b). All three species of *Squatina* from the eastern Atlantic are Critically Endangered based on former abundance, severe fishery-driven declines and rarity of recent records, and quantifying landings in West Africa is a conservation priority given minimal current knowledge (Morey et al. 2007a, b; Gordon et al. 2017). Diop and Dossa (2011) recorded *S. oculata* and *S. aculeata* as 'frequent' and 'quite frequent' respectively in mainland SFRC countries, but this is in marked contrast to the narrative from other data sources (e.g. from Russian trawl surveys and Senegalese artisanal fishers) which reported both species as being common around the 1970 s and 1980 s, but as extremely rare now. Regional research surveys recorded very sparse occurrence, with the last Gambian records for *S. oculata* consisting of six individuals between 1986 to 2000, and of a single *S. aculeata* in 1998 (Morey et al. 2007a, b). The removal of fins recorded in our surveys is of interest given that the shark fin trade has not previously been identified as a specific threat to *Squatina* species (Clarke et al. 2007; Morey et al. 2007a, b; Gordon et al. 2017).

Endangered batoid species

The common guitarfish *Rhinobatos rhinobatos* (Rhinobatidae) was never recorded in abundance, although single individuals or small numbers were frequently recorded at landing and processing sites in all years, either as fresh landings or on drying racks (Table 1). While fins were removed by fishers on some fresh specimens (including small individuals of ca. 50 cm TL) this practice did not appear to be consistent (or as consistent as that observed for *G. cemiculus*) as intact fins were sometimes observed on drying whole specimens. *Rhinobatos rhinobatos* was previously described as abundant and one of the most common elasmobranchs in the region, frequently or very frequently captured in the Gambia

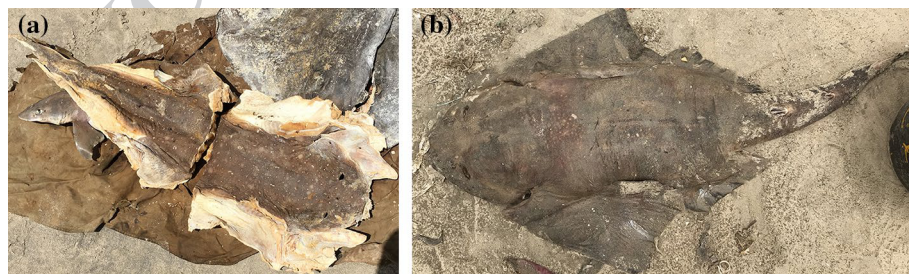


Fig. 6 Critically Endangered angel sharks recorded in The Gambia. **a** Smoothback angel shark *Squatina oculata*, estimated 100 cm TL, 5th April 2017 (credit: Kayley Knight) **b** sawback angelshark *Squatina aculeata*, estimated 150 cm TL, 25th March 2018 (credit: Jess Holdsworth)

and all other mainland SFRC countries (Notarbartolo di Sciara et al. 2007; Diop and Dossa 2011; Séret 2016a). Analysis of trawl data from Mauritania revealed strong declines in *R. rhinobatos* abundance over the period 1990–2010 (Meissa and Gascuel 2015). The global assessment for *Rhinobatos rhinobatos* is Endangered based on past and suspected future declines, such as evidence of regional extinction in the northern Mediterranean (Notarbartolo di Sciara et al. 2007).

The medium-sized daisy whiplay *Fontitrygon margarita* was only recorded in small numbers on two occasions (Table 1) at Ghanatown amongst fresh landings dominated by *G. cemiculus*. Regional surveys by Diop and Dossa (2011) only recorded this species in Guinea where it was ‘very rare’, and it is currently considered Endangered (Compagno and Marshall 2016, based on a 2004 assessment) given larger size, low fecundity and apparent declines in abundance.

Species of exceptional evolutionary distinctness

Individuals of the striped panray *Zanobatus schoenleinii* (Zanobatidae) were frequently observed (at least in 2011, 2014, 2015 and 2018) at landing sites such as Gunjur, often in some abundance, consistent with surveys of the wider region (Diop and Dossa 2011, Séret 2016c). Whole fresh or dessicated individuals were always found discarded on the ground around fishing communities indicating it is not consumed or used in The Gambia, which contrasts with reported utilization in Guinea (Valenti 2009). This species has recently been identified as the single most evolutionary distinct species of all chondrichthyan fishes (Stein et al. 2018), and is currently assessed as Data Deficient (Valenti 2009). A second species of the genus *Zanobatus*, *Z. maculatus*, was recently described from the Gulf of Guinea (Séret 2016d), although its true distribution in West Africa needs to be determined.

Rarely recorded endemic species

Two electric rays (Torpedinidae) endemic to West Africa were found discarded whole on the beach, comprising one individual each of the rosette torpedo *Torpedo bauchotae* and West African torpedo *T. mackayana* (Table 1). Both species are Data Deficient and rarely recorded, with *T. bauchotae* known from only a few specimens (Pheeha and Cronin 2009; Séret et al. 2009; Dossa and Diop 2011; de Carvalho et al. 2016). In addition, two individuals of a scyliorhinid shark assumed to be the regionally endemic West African catshark *Scyliorhinus cervigoni* (Table 1) had been skinned and had fins drying on nearby racks. This species is very poorly known and was either absent or very rare in previous surveys (Burgess 2006; Diop and Dossa 2011).

Data deficient species

A suite of at least four Data Deficient stingrays (Dasyatidae) recorded occasionally or rarely (whole and fresh or dried) comprised brown stingray *Bathytoshia lata*, smalltooth stingray *Hypanus rudis*, marbled stingray *Dasyatis marmorata* (Notarbartolo di Sciara et al. 2009), common stingray *D. pastinaca* (Serena et al. 2009a) and round stingray *Taeniurops grabatus* (Serena et al. 2009b) (Table 1).

While *B. lata* is currently considered Least Concern, the 2007 assessment predates taxonomic revision and therefore requires updating; Eastern Atlantic populations of *B. lata*

278 were known as *Dasyatis centroura* (Ebert et al. 2016; Last et al. 2016b). Previous surveys
279 of the region only recorded *D. centroura* in Guinea, as ‘very rare’ (Diop and Dossa, 2011),
280 possibly due to difficulties discriminating this species. The taxonomic status of *Hypanus*
281 *rudis* is uncertain, and our records are of note as this species is known only from the origi-
282 nal description from the Nigerian coast, with the holotype probably lost and no records
283 since (Last et al. 2016b; Séret et al. 2016b). Separate photographic records of very large
284 stingrays (>200 cm DW) obtained locally may be *H. rudis* (Ruth H. Leeney, unpublished
285 data). Only two individuals of the distinctive *D. marmorata* were recorded which is con-
286 sistent with rarity in surveys of the region by Diop and Dossa (2011), yet it was apparently
287 not uncommon off Mauritania in 1998–2002 where over 1000 individuals were recorded
288 in surveys of fish landing sites (Valadou et al. 2006). *Dasyatis pastinaca* is difficult to dis-
289 criminate from the reportedly sympatric *D. tortonesei* (Last et al. 2016b), and there may
290 be further cryptic diversity within *D. pastinaca* (BS pers. obs.). In previous surveys of the
291 region *D. pastinaca* was reported as uncommon, and not recorded from Senegal or Gambia
292 (Diop and Dossa 2011). While the global Red List Assessment for *D. pastinaca* is Data
293 Deficient (Serena et al. 2009a), a more recent assessment for the Mediterranean was Vul-
294 nerable, based on suspected declines (Serena et al. 2016). A single *Taeniurops grabatus*
295 recorded in our survey reflects an absence of mainland records in previous regional surveys
296 (Diop and Dossa 2011).

297 Only two individuals of the bullray *Aetomylaeus bovinus* were recorded (Table 1). In
298 1989 this species (reported as *Pteromylaeus bovinus*) was one of the most common elas-
299 mobranchs in research fishing off Guinea-Bissau by a Portuguese institute, and in 2002
300 it was common off Senegal (Diop and Dossa 2011; Wintner 2016). However it was not
301 recorded at all in the region in more recent surveys, excepting in Guinea where it was ‘very
302 rare’. The current global assessment for this species is Data Deficient, but is considered
303 Critically Endangered in the Mediterranean where severe declines are suspected (Walls and
304 Buscher 2016; Wintner 2016).

305 Not evaluated species

306 At least four taxa were recorded that have not been evaluated by the IUCN *Red List*, con-
307 sisting of species that have only recently been described or are undergoing taxonomic treat-
308 ment. Dogfish sharks (Squalidae) were recorded on one occasion in 2011 (Table 1) but
309 could not be confidently identified to species level as most individuals had pectoral, dor-
310 sal, and lower and upper caudal fin lobes removed. The taxonomic status of *Squalus* in
311 the eastern Atlantic is highly complex (Verissimo et al. 2016) with revision ongoing; sev-
312 eral species could potentially occur off West Africa, including the newly described Smith’s
313 dogfish shark *Squalus margaretsmithae* (Viana et al. 2017) and *S. probatovi*, a valid spe-
314 cies known from Angola that is to be resurrected (Sarah Viana, pers. comm.; Viana and
315 Carvalho 2018).

316 Two species within the cryptic brown skate *Raja miraletus* species complex were occa-
317 sionally recorded: the newly described *R. parva*, and the undescribed *R. cf. miraletus* (Last
318 and Séret 2016) (Table 1). The apparent rarity of species in this complex in our surveys
319 is notably different to it (as *R. miraletus*) being reported as ‘very frequent’ and one of the
320 commonest elasmobranchs of the region (Diop and Dossa 2011).

321 As noted, most processed specimens of butterfly rays (Gymnuridae) were only identi-
322 fied to genus (*Gymnura* sp.), although some fresh individuals were identified as the newly
323 described *G. sereti* (Yokota and de Carvalho 2017). A larger species, *G. altavela* is also

known to occur on the coast of West Africa (Yokota et al. 2016) and larger individuals we recorded are likely to have been this species (Table 1). While our surveys recorded *Gymnura* in most years they were only ever present as one, or a few (<10), individuals fresh or on drying racks and assumed to be consumed for food. In previous surveys of the region, Diop and Dossa (2011) only recorded *Gymnura* in Guinea where they were 'very rare'. While the global assessment for *G. altavela* is Vulnerable (Vooren et al. 2007), Mediterranean populations are Critically Endangered based on large size, disappearance from historic range and suspected significant declines in abundance (Walls et al. 2016).

Cownose rays (Rhinopteridae) were also only identified to genus (*Rhinoptera*) as cryptic diversity exists in West Africa, with at least one undescribed species possibly present (Last et al. 2016c). *Rhinoptera* were only recorded occasionally as single or a few individuals in most years, but in 2015 a total of 961 individuals were recorded on drying racks at Gunjur and Ghanatown (Fig. 7). Based on size at maturity (LT_{50}) of 77.5 cm and 80.2 cm DW for male and female *R. marginata* respectively from Mauritania (Valadou et al. 2006) only 1% or less of this catch was of a size likely to be mature. Significant conservation concern has been raised for *Rhinoptera* species due to exceptionally low fecundity of one pup and susceptibility to intensive fisheries from aggregation in large schools. A congener in the western Atlantic has reportedly been extirpated from southern Brazil due to fisheries (Vooren and Lamónaca 2004).

Other shark species

Milk shark *Rhizoprionodon acutus* were occasionally recorded as single or a few individuals in all years, with a single photograph of numerous individuals on drying racks from the 2010 survey possibly indicating greater abundance in the wet season. It was recorded in abundance in The Gambia in 2010–2011 (Bojang 2011); this species is considered Least Concern (Simpfendorfer 2003). The remaining shark species were recorded occasionally or rarely (Table 1) and included young scalloped hammerhead *Sphyrna lewini*, which is Endangered (Baum et al. 2007). Other shark taxa recorded were smoothhound *Mustelus mustelus* (Vulnerable, Serena et al. 2009c), and barbelled houndshark *Leptocharias smithii*

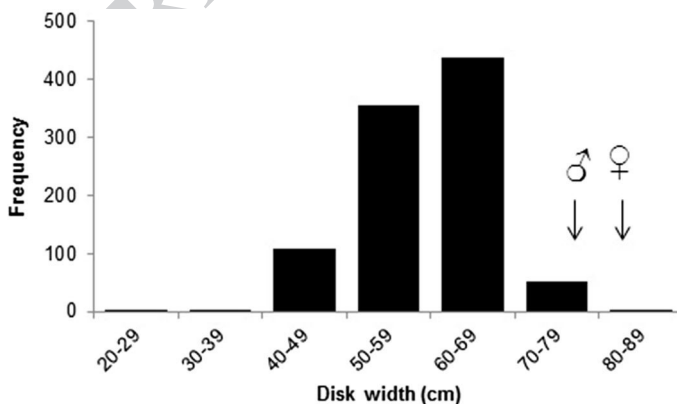


Fig. 7 Disk width-frequency distribution of 961 cownose ray *Rhinoptera* sp., Ghanatown and Gunjur, The Gambia, April 2015. Arrows indicate approximate size at 50% maturity for males and females of *Rhinoptera bonasus* from Mauritania (Valadou et al. 2006)

and bull shark *Carchahinus leucas*, both of which are Near Threatened (Compagno 2005; Simpfendorfer and Burgess 2009).

Value of the study approach

The core of this survey was essentially the collation of many hundreds of photographs by undergraduate students with no previous experience of elasmobranchs, yet it has provided important insights into several poorly known aspects of their diversity and fisheries in West Africa. These include (1) expanding knowledge on diversity, biology and species composition, including new data on size (*F. margaritella*) and notable records of Critically Endangered, rare and poorly known species (e.g. *Squatina* spp., *Hypanus rudis*); (2) documenting biological indicators of the fishery such as relative abundance and, most notably, immature juveniles forming a significant component of catches of key commercial species (*G. cemiculus*, *F. margaritella*, *Rhinoptera* sp.); (3) documenting fisheries and utilisation, including apparently targeted fisheries for high-value species (*G. cemiculus*), significant and previously undocumented fisheries mortality for small species, both for consumption (*F. margaritella*) and as discards (*Z. schoenleinii*), and processing of shark fins in taxa not generally considered important in the fin trade (*Squalus* spp., *Squatina* spp., *Scyliorhinus cervigoni*); and (4) providing evidence of the apparent importance of specialised elasmobranch industries to coastal livelihoods in The Gambia. Despite its largely qualitative and low-resolution nature, the study has clearly demonstrated a valuable and cost-effective approach to elasmobranch data collection in the developing world, and more broadly helped aid the call for more data on artisanal fisheries in the region (Belhabib et al. 2018).

Rarity and absence

Our surveys sampled an intensive shallow water demersal fishery, which in some cases targeted shark-like batoids (guitarfishes), and was dominated by batoids. Several taxa were recorded in our surveys only rarely, and several species that might have been expected to occur did not. This rarity is likely due to reflect the influence of many factors including bias from sampling and fisheries practices, and, in some cases, true abundance. The most significant influence is likely to be due to nearly all our surveys being limited to March and April, under-sampling taxa present in other months. Further, as the fishery appeared to operate mostly in shallow waters it is unlikely to effectively sample taxa generally found in deeper, cooler water in the tropics (e.g. *Squalus* and *Raja*), and at-sea discarding practices of undesirable species (e.g. *Torpedo*) may also have contributed to rarity of records. Some rarity might also be explained by our study location being towards the edge of a species' known distribution (e.g. *T. bauchotae*, *F. margarita*).

Even taking these factors into account, a number of demersal elasmobranch species that we recorded only rarely have been (1) previously reported as abundant in the region (2) have shown evidence of fishery-driven declines elsewhere in their range and (3) would be expected to be caught in the fishery; examples include angel sharks (*Squatina* sp.) and bull-ray *Aetomylaeus bovinus*. Similarly, it was notable that over 7 years of survey we failed to record evidence of a single individual of a number of highly threatened species of large sharks and shark-like batoids whose distribution includes The Gambia that are known or suspected to have experienced severe declines (e.g. great hammerhead *Sphyrna mokarran*, sawfishes *Pristis* sp., African wedgefish *Rhynchobatus luebberti*) (Robillard and Séret 2006; Denham et al. 2007; Diop and Dossa 2011; Leeney and Downing 2016; Moore 2017).

396 Risks to biodiversity, coastal livelihoods and food security

397 In addition to those may have already been severely depleted, it is highly likely that the
398 trajectory of relatively abundant species we recorded will follow a similar path in the near
399 future due to their limiting life history characters and intensive artisanal fisheries. As noted
400 above the large-bodied *G. cemiculus* is known to have declined elsewhere in its range,
401 reflecting reported declines of larger guitarfishes (Moore 2017). The intensive utilisation of
402 the small and apparently previously unused stingray *F. margaritella* (for which little or no
403 biological information exists on which to base extinction risk assessment) may represent
404 a case of fishing down marine food webs, and indicate that exploitation is unsustainable
405 (Pauly et al. 1998). Indeed, some fishers we spoke with indicated the batoid fishery had
406 developed in response to a decline of the shark fishery. As well as a depletion of elasmobranch
407 abundance, the fishery represents a risk to taxa that are almost entirely unknown,
408 including potentially undescribed species. Collapse of regional elasmobranch fisheries
409 could have far-reaching biodiversity consequences; Brashares et al. (2004) demonstrated
410 increased hunting of terrestrial wildlife for bushmeat in years of poor fish supply in West
411 Africa.

412 Our study identified fisheries and industries that appear to be wholly or largely reliant
413 on elasmobranchs, although further research on socio-economic aspects is clearly needed.
414 Given the known importance of fishing to dietary protein, income and employment in
415 The Gambia and its limited and vulnerable food production options (Moseley et al. 2010;
416 United Nations 2014; FAO 2018a), further depletions or collapse of elasmobranch popu-
417 lations could have far-reaching consequences for humans. Overexploitation and declines
418 of sharks have been highlighted as one of the factors in a potential food security crisis in
419 Madagascar (Le Manach et al. 2012).

420 Challenges and solutions

421 West African fisheries in general face multiple significant challenges, including strong
422 declines in artisanal catch per unit effort (Belhabib et al. 2018), intensive and often unre-
423 ported overfishing by foreign interests including the EU and China (Pala 2013; Ramos and
424 Grémillet 2013), open access waters intensively exploited by migrant fishers (Binet et al.
425 2012) and predicted significant decline in fisheries production from climate change (Lam
426 et al. 2012). In this context, solutions to unsustainable elasmobranch fisheries may not
427 seem attainable. However, in The Gambia elasmobranch fisheries may be being supported
428 by a cycle of loans and debt, with fishers apparently willing to seek alternative livelihoods
429 if capital to start other businesses was available (Mendy 2011); this may present opportuni-
430 ties for conservation intervention. The smoking of small pelagic fishes for export has also
431 been proposed as an alternative livelihood (Diop and Dossa 2011); but this comes with its
432 own conservation and sustainability issues including use of mangrove wood and the over-
433 exploited status of small pelagics (Diop and Dossa 2011; Polidoro 2016).

434 One possible solution to help alleviate risks to both elasmobranchs and human com-
435 munities is the development of accredited fisheries. The current bycatch of guitarfishes
436 and other threatened elasmobranchs may represent a constraint to social and economic
437 development in The Gambia by preventing access to high-value eco-labelled export
438 markets. The bottom gillnet sole fishery is of major interest to EU markets and retail-
439 ers, but has previously failed Marine Stewardship Council sustainability accreditation

pre-assessment on aspects including a lack of catch-specific information on bycatch species (including 'high risk' guitarfish) (Gabis et al. 2012; Marine Stewardship Council 2013; Coastal Resources Centre 2014). The current paper goes some way to providing these missing data.

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